

## Integrating the Water Framework Directive into the Habitats Directive: Analysis of distribution patterns of lacustrine EU habitats in lakes of Lombardy (northern Italy)

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### ABSTRACT

The existence of strong potential synergies between the Water Framework Directive (WFD) and the Habitats Directive (HD) is widely acknowledged. Indeed, ensuring favourable conservation conditions for aquatic habitats and species of conservation concern is closely related to the achievement of a good ecological status in water bodies. However, since these two sets of European laws are generally applied without any coordination, an inefficient use of resources may adversely affect their goals. The main negative outcome is an increase in the cost of monitoring programs for collecting data in nature (*i.e.*, physical and chemical parameters, species and habitats, plant communities). The use of macrophytes as a bioindicator, as imposed by the WFD, may instead help to integrate data on aquatic EU habitats and enhance knowledge of such habitats outside the Natura 2000 network. The aim of present study was to evaluate the usefulness of data collected in WFD monitoring surveys as a means of inferring the occurrence and the distribution of lacustrine aquatic habitats in countries belonging to the European Union (EU). The main aim of the analysis was to identify the depth gradient distribution of diagnostic macrophyte *taxa* in two EU habitats (3140, *i.e.*, *Chara*-dominated benthic communities, and 3150, *i.e.*, natural eutrophic lakes) using data collected in lakes in Lombardy (northern Italy), some of which are included in the Natura 2000 network (10 out 16). While recognizing the limitations of the data collected within the two frameworks, the results confirmed the marked usefulness of WFD data as a means of enhancing the knowledge available on lacustrine aquatic habitats in the EU. WFD data can actively help to improve the basic information on aquatic habitats, thereby more effectively supporting regional strategies for biodiversity conservation as well as recovery programs.

**Key word:** WFD; HD; monitoring programs; aquatic vegetation; vascular macrophytes; charophytes; nutrient drivers; macrophyte depth distribution.

### INTRODUCTION

The Water Framework Directive (WFD, European Union, 2000) and the Habitats Directive (HD, European Union, 1992) are strongly interrelated (Janauer *et al.*, 2015). The synergy between these two directives represents an intriguing goal aimed at pursuing an effective conservation of aquatic ecosystems. It being universally accepted that the achievement of favourable conservation conditions for aquatic habitats or species of conservation concern is closely related to the possibility of guaranteeing a good ecological status of colonized water bodies (Janauer *et al.*, 2015). However, each of these directives is currently implemented on its own, thereby potentially hampering the achievement of their respective goals (Benun *et al.*, 2009). One major effect of the lack of coordination between the directives is an increase in the cost of monitoring programs for collecting data in nature (*i.e.*, physical and chemical parameters, species and habitats, plant communities). Indeed, the synergy and potential

conflicts between these directives have mainly been investigated in terms of the attainment of their respective goals rather than of the possible reciprocal operational support (Ecke *et al.*, 2010; Collins *et al.*, 2012; Janauer *et al.*, 2015).

WFD monitoring activities include both biotic elements and physical and chemical features (European Union, 2000), providing the opportunity to integrate the structural/compositional investigation of biotic communities with their environmental drivers (*e.g.*, hydro-morphological, physical and chemical determinants). Indeed, the depth distribution of lacustrine macrophytes is affected by several ecological factors: light (Canfield *et al.*, 1985; Chambers and Kalff, 1985), phosphorus (Søndergaard *et al.*, 2013), temperature and depth of the thermocline (Genkai-Kato and Carpenter, 2005), and pressure (Dale, 1986). Although light seems to be the most representative ecological factor, phosphorus availability is also important on account of the role it plays in the regulation of macrophyte distribution in a lake as well as of their occurrence

at the regional scale (Duarte and Kalff, 1990). Furthermore, the maximum depth of macrophyte colonization ( $Z_{c_{max}}$ ) is strongly dependent on the size of the lake, which is in turn directly regulates the depth of the thermocline (Genkai-Kato and Carpenter, 2005). Where water transparency is high, the  $Z_{c_{max}}$  is greater in large lakes than in medium-sized and small lakes (Azzella *et al.*, 2014a, 2014b). These factors may be useful for monitoring the conservation status of aquatic European Union (EU) habitats if we bear in mind that the HD reporting activities are – *inter alia* – based on an evaluation of the area and range occupied by habitats (Evans and Arvela, 2011).

The level of information available for aquatic vegetation is generally far lower than that available for terrestrial vegetation. A greater integration between these two directives may represent a win-win strategy to fill the gaps in knowledge on aquatic vegetation and to assess the effectiveness of WFD legislation in improving the quality of water bodies. This is particularly interesting if we consider that aquatic ecosystems are among the ecosystems threatened most on a worldwide scale and that, besides the inherent difficulties associated with sampling, the cost of providing reliable data is high (*e.g.*, boat availability, poor weather conditions, wind) (Azzella *et al.*, 2013a).

A potential change in this scenario may lie in the WFD. It provides for the use of several biotic communities as a means of assessing the quality status of water bodies, including macrophytes, fish and macroinvertebrates, which are actually the same target elements as those found in the HD (Janauer *et al.*, 2015). By focusing on lacustrine macrophytes, we may be able to integrate the HD database with routine monitoring information

yielded by the WFD. Indeed, almost all the aquatic vegetation in lakes, whether it is dominated by vascular plants or charophytes, can be referred to the conservation habitats defined by the HD (European Union, 1992; Bolpagni, 2013a).

The main aim of the present study was to verify the occurrence of the EU habitat codes 3140 and 3150 by using WFD data, both within and outside the Natura 2000 network in Lombardy. By considering the depth distribution of their diagnostic *taxa*, we also investigated the predictive role of water transparency (expressed as SD) in modelling the depth distribution of lacustrine EU habitats.

## METHODS

### Study area

The study was conducted on data collected from 16 different lakes in the Lombardy Region (northern Italy), 10 of which are included in the Natura 2000 network (Tab. 1; Fig. 1). The lakes covered a wide range of ecological conditions (Tab. 2), both as regards their physical (*e.g.*, depth) and trophic (*e.g.*, water transparency investigated by means of Secchi disk, SD) characteristics. Lake depth ranged from 3 m (Lake Ganna) to 370 m (Lake Maggiore), whereas the SD ranged from ~0.80-1.00 m (lakes of Mantova) to 8.42 m (Lake Monate).

### Macrophyte EU target habitats

The target habitats of the present study are the EU habitats 3140 (hard oligo-mesotrophic waters with benthic vegetation of *Chara* spp.) and 3150 (natural eutrophic

**Tab. 1.** Interactions between lakes and Natura 2000 network; the official name of the site, its national code and type, as well as the percentage of the lacustrine area included in the site is reported for each lake included in a Natura 2000 site. The overall assessment of EU habitat 3150 is also included (updated in February 2016).

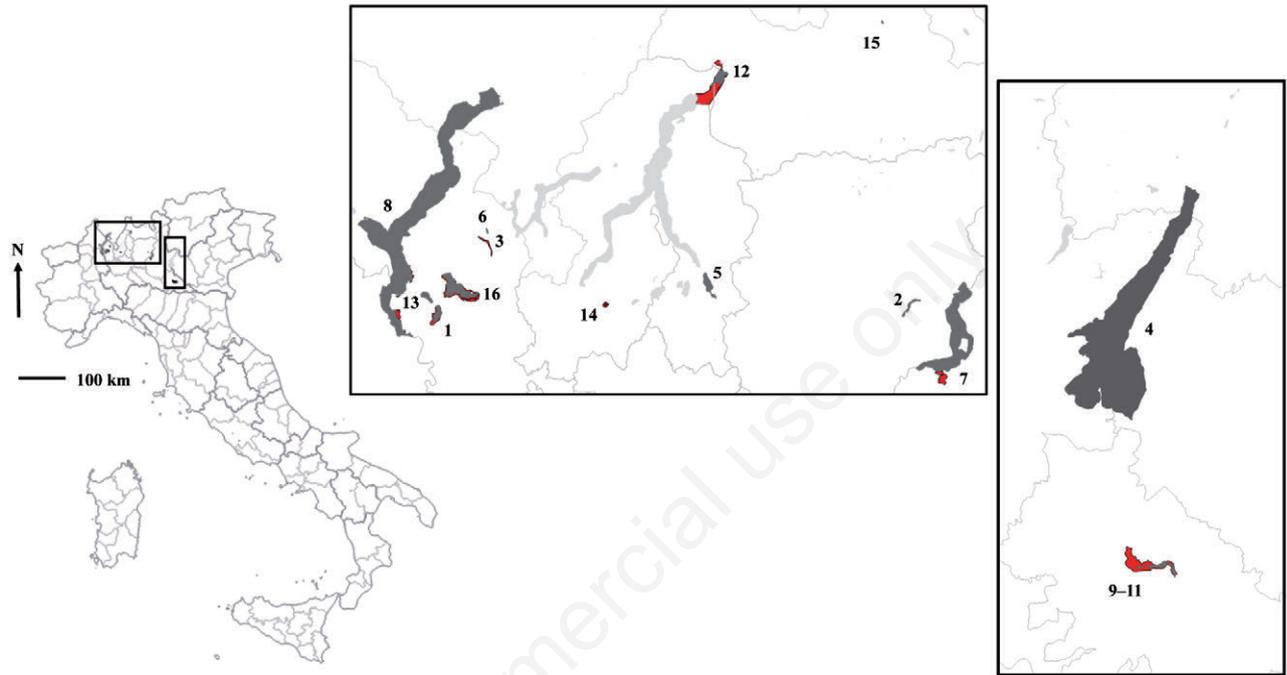
Lake	Natura 2000 site name	National code	Type	LA	3150
Comabbio	Lago di Comabbio	IT2010008	SAC	100%	B
Ganna	Lago di Ganna	IT2010001	SAC	100%	
Iseo	Torbiera d'Iseo	IT2070020	SAC/SPA	<1%	B
Maggiore	Canneti del Lago Maggiore	IT2010502	SPA	<1%	B
	Palude Bruschera	IT2010015	SAC	<1%	B
Mantova Superior	Ansa e Valli del Mincio	IT20B0017	SAC	40%	C
	Valli del Mincio	IT20B0009	SPA	100%	C
Mantova Middle	Valli del Mincio	IT20B0009	SPA	100%	C
Mantova Inferior	Valli del Mincio	IT20B0009	SPA	100%	C
Mezzola	Lago di Mezzola e Pian di Spagna	IT2040042	SAC	100%	A
	Lago di Mezzola e Pian di Spagna	IT2040022	SPA	100%	A
Montorfano	Lago di Montorfano	IT2020004	SAC	100%	
Varese	Alnete del Lago di Varese	IT2010022	SAC	<1%	B
	Lago di Varese	IT2010501	SPA	100%	B

LA, lacustrine area; SAC, Special Area of Conservation; SPA, Special Protection Areas; A, excellent value; B, good value; C, significant value.

lakes with *Magnopotamion* or *Hydrocharition*-type vegetation). The former includes oligo- to meso-trophic waters with well-developed benthic vegetation dominated by stoneworts (charophytes); the latter includes the free-floating and rhizophytic plant communities of eutrophic

lakes (Biondi *et al.*, 2009; Azzella *et al.*, 2013b; Bolpagni, 2013a).

Generally, *Chara*-dominated vegetation has a low species richness and is strictly controlled by water nutrient content (especially by phosphorous availability) (Blindow,



**Fig. 1.** Study area; the spatial distribution of the lakes analyzed (in dark grey), and the Natura 2000 sites (in red) are reported; see Tab. 1 and Tab. 2 for the Natura 2000 site codes and the lake names, respectively.

**Tab. 2.** Lakes features and sampling effort for each investigated lake.

Lake	Alt (m asl)	A (km <sup>2</sup> )	D (m)	V (m <sup>3</sup> 10 <sup>6</sup> )	SD (m)	TP (μg L <sup>-1</sup> )	IY	#TR	#SP
1 Comabbio	243	3.6	8	16.6	3.23	5.3	2008	23	78
2 Endine	334	2.1	9	11.9	2.42	19.0	2008	67	157
3 Ganna	390	0.1	3	0.1	2.50	2.5	2008	5	15
4 Garda	65	368.0	350	49031.0	8.20	17.7	2011	82	760
5 Garlate	205	4.6	34	70.0	5.62	23.0	2011	20	164
6 Ghirla	415	3.2	14	nd	4.93	2.5	2008	18	74
7 Iseo	185	61.0	251	7600.0	4.67	61.5	2008	109	743
8 Maggiore	193	213.0	370	37500.0	6.42	4.4	2012	198	1056
9 Mantova Middle	15	3.7	12	nd	0.85	50.0	2010	28	65
10 Mantova Inferior	15	1.5	9	nd	0.82	37.5	2010	20	51
11 Mantova Superior	18	1.1	15	nd	0.99	55.0	2010	45	119
12 Mezzola	200	5.9	69	149.0	2.72	11.3	2011	30	189
13 Monate	266	2.5	34	45.0	8.42	2.5	2008	40	172
14 Montorfano	397	0.5	7	1.9	2.83	20.5	2012	12	48
15 Palù	1921	0.2	50	nd	6.43	5.0	2011	10	52
16 Varese	238	14.8	26	160.0	3.38	39.3	2008	44	130

Alt, altitude; A, area; D, depth; V, volume; SD, Secchi disk; TP, total phosphorous; IY, investigation year; #TR, total number of transects; #SP, sampling plots; nd, not detected.

1992). Since the *Chara*-dominated communities in lakes in northern Italy have seldom been investigated, few data are available on their composition and conservation status (Bolpagni *et al.*, 2013). By contrast, more detailed investigations have been conducted on the volcanic lakes of central Italy (Azzella *et al.*, 2013a, b, 2014; Bolpagni *et al.*, 2016). The free-floating and rhizophytic plant communities are found both in eutrophic lakes and ponds. Almost all aquatic plant communities can be referred to this habitat type (*e.g.*, *Lemnetea* and *Potametea* vegetation classes) (Biondi *et al.*, 2009), including free-floating [*e.g.*, *Lemna* ssp. or *Salvinia natans* (L.) All. dominated stands] and rooted vegetation consisting of both submerged (*e.g.*, *Vallisneria natans* L., *Potamogeton lucens* L. dominated stands) and emergent [*e.g.*, *Nuphar lutea* (L.) Sm., *Nymphaea alba* L., *Potamogeton natans* L. dominated stands] plant communities (Bolpagni and Piotti, 2015, 2016).

No record of the presence of EU habitat 3140 is present in the standard data forms of the Natura 2000 sites, whereas the presence of EU habitat 3150 is recorded in 8 out of 10 lakes (lakes Comabbio, Iseo, Maggiore, Mantova Superior, Middle and Inferior, Mezzola, and Varese)

(Tab. 1). Furthermore, no standardized information on any EU habitats is available for the lakes outside the Natura 2000 sites (lakes Endine, Garda, Garlate, Ghirla, Monate, and Palù). Thus, the data currently available indicate that EU habitat 3140 is not present, whereas EU habitat 3150 covers ~298 ha.

### Study design and data analysis

We analyzed data collected in WFD monitoring surveys conducted between 2009 and 2012. Data included SD and TP values (Tab. 2) and the depth of macrophyte distribution. Macrophyte data was collected by applying the Italian national protocol, as described by Oggioni *et al.* (2011). Accordingly, the lakeshores were divided into homogeneous sections by inspecting helophytic and macrophyte vegetation. Data on macrophytes were recorded from a depth of 0.5 m to the maximum depth of colonization at 1-meter depth intervals along transects in a randomly selected section (Bolpagni, 2013a,b). The cover-abundance – expressed as a percentage – was estimated for all the *taxa* identified.

Three different morpho-functional groups of *taxa* were identified (Tab. 3) on the basis of the diagnostic

**Tab. 3.** Total representativeness of diagnostic *taxa* (total sampling plots colonized) of the EU habitat codes 3150 (natural eutrophic lakes; 1: free-floating and floating-leaved rhizophyte-dominated stands, and 2: submerged plant-dominated stands) and 3140 (3: charophyte-dominated benthic communities).

Habitat code		Species	Total
3150	1	<i>Nuphar lutea</i> (L.) Sm.	97
3150	1	<i>Nymphaea alba</i> L.	312
3150	1	<i>Nymphoides peltata</i> (S.G. Gmel.) Kuntze	40
3150	1	<i>Trapa natans</i> L.	241
3150	2	<i>Ceratophyllum demersum</i> L.	1651
3150	2	<i>Myriophyllum spicatum</i> L.	1683
3150	2	<i>Najas marina</i> L.	1849
3150	2	<i>Najas minor</i> All.	73
3150	2	<i>Potamogeton lucens</i> L.	31
3150	2	<i>Potamogeton pectinatus</i> L.	297
3150	2	<i>Potamogeton perfoliatus</i> L.	916
3150	2	<i>Potamogeton pusillus</i> L.	274
3150	2	<i>Potamogeton trichoides</i> Cham. & Schldl.	20
3150	2	<i>Vallisneria spiralis</i> L.	2809
3150	2	<i>Zanichellia palustris</i> L.	582
3140	3	<i>Chara globularis</i> Thuiller	1439
3140	3	<i>Chara intermedia</i> A.Braun	36
3140	3	<i>Chara tomentosa</i> Linnaeus	25
3140	3	<i>Nitella flexilis</i> (Linnaeus) C.Agardh	61
3140	3	<i>Nitella gracilis</i> (J.E.Smith) C.Agardh	27
3140	3	<i>Nitella hyalina</i> (De Candolle) C.Agardh	36
3140	3	<i>Nitellopsis obtusa</i> (N.A.Desvieux) J.Groves	84

power of each species detected in terms of life-growth form and the phytosociological literature. The free-floating and floating-leaved rhizophytes (group 1) and submerged rhizophytes (2) were referred to EU habitat 3150; charophytes (3) were instead referred to EU habitat 3140. *Taxa* with at least 20 records (=22) were considered in the analysis (Supplementary Tab. 1). Overall, 751 transects were analyzed, which resulted in a total of 3,873 plots being investigated and 14,130 *taxon* records being collected. We considered a habitat present when at least one stonewort or vascular diagnostic species was present with an estimated cover-abundance higher than 35%.

Simple regression analysis was used to evaluate the relationships between the depth distribution of the EU habitats being analyzed and SD as a proxy of the underwater light conditions. The area and depth descriptors were excluded from the analyses owing to their high collinearity with SD. All the analyses were performed in the R environment (R Development Core Team, 2016), considering the depth distribution of the EU habitats' diagnostic *taxa*, considering the three morpho-functional groups of species separately: i) free-floating and floating-leaved plants, ii) submerged plants, and iii) charophytes.

## RESULTS

### Macrophyte EU habitat distribution patterns

Based on the diagnostic species spatial arrangement, we confirmed the presence of EU habitat 3150 in all the lakes investigated; by contrast, EU habitat 3140 was present in only 7 of the 16 lakes (43.8%) (Fig. 2).

The emergent plant communities of EU habitat 3150 grew to a maximum depth of 4.5 m (recorded at Lake Monate), with a mean colonization depth of  $1.3 \pm 0.8$  m ( $\pm$  SD) (Fig. 2), whilst the submerged EU 3150 plant communities grew down to 10.5 m (recorded at Lake Garda), with a mean colonization depth of  $2.9 \pm 1.9$  m (Fig. 2). The emergent EU 3150 communities were dominated by *N. lutea*, *N. alba*, *Nymphaoides peltata* (S.G. Gmel.) Kuntze, and *Trapa natans* L., the submerged communities by *Ceratophyllum demersum* L., *Myriophyllum spicatum* L., *Najas marina* L., *Potamogeton lucens* L., *P. pectinatus* L., *P. perfoliatus* L., and *Vallisneria spiralis* L.

EU habitat 3140 exhibited a greater variability in terms of colonized depths than the vascular plant communities, with a maximum depth of 16.5 m at Lake Garda, and a mean colonization depth of  $3.8 \pm 2.6$  m (Fig. 2). The prevalent (dominant) species were *Chara globularis* Thuiller, *C. intermedia* A.Braun, *C. tomentosa* Linnaeus, *Nitella flexilis* (Linnaeus) C.Agardh, and *Nitellopsis obtusa* (N.A.Desvaux) J.Groves.

### Depth distribution of lacustrine EU habitats vs SD

The linear regression analysis revealed that the mean depth range colonized by submerged vegetation correlated with SD (Fig. 3). Both submerged EU 3150 and *Chara*-dominated communities positively correlated with SD ( $R^2$  adj=0.501,  $P < 0.01$ ,  $n=16$ , and  $R^2$  adj=0.737,  $P < 0.01$ ,  $n=7$ , respectively); by contrast, no correlation was detected between emergent plant communities (EU 3150) and SD ( $R^2$  adj=0.138,  $P=0.864$ ,  $n=9$ ).

## DISCUSSION

The present data confirm that information collected during WFD monitoring programs are able to make an important contribution to our knowledge of aquatic EU habitats in lacustrine environments. These results may, in addition, be used to evaluate the effectiveness of local water conservation policies. For example, the fact that EU habitat 3140 had not previously been found in the Natura 2000 sites considered in this study may be due to a progressive improvement in the chemical and physical conditions of lakes. As is widely known, *Chara*-dominated communities are considered to be among the aquatic habitats that are most sensitive to external pressures, especially in terms of nutrient loading (e.g., phosphorous availability or algal blooms) (Blindow, 1992; Auderset Joye *et al.*, 2002). We cannot, however, exclude inaccuracies in the assessment of the structure and composition of the submerged habitats during the past routine HD surveys given the intrinsic difficulties encountered when monitoring water bodies (Azzella *et al.*, 2013b), as well as the rarity or temporal dynamicity of these communities (Bolpagni *et al.*, 2016). Nevertheless, these uncertainties further support the pivotal contribution of the use of the standardized approaches adopted within the framework of the WFD to integrate current knowledge on the presence, areal distribution and representativeness of aquatic EU habitats.

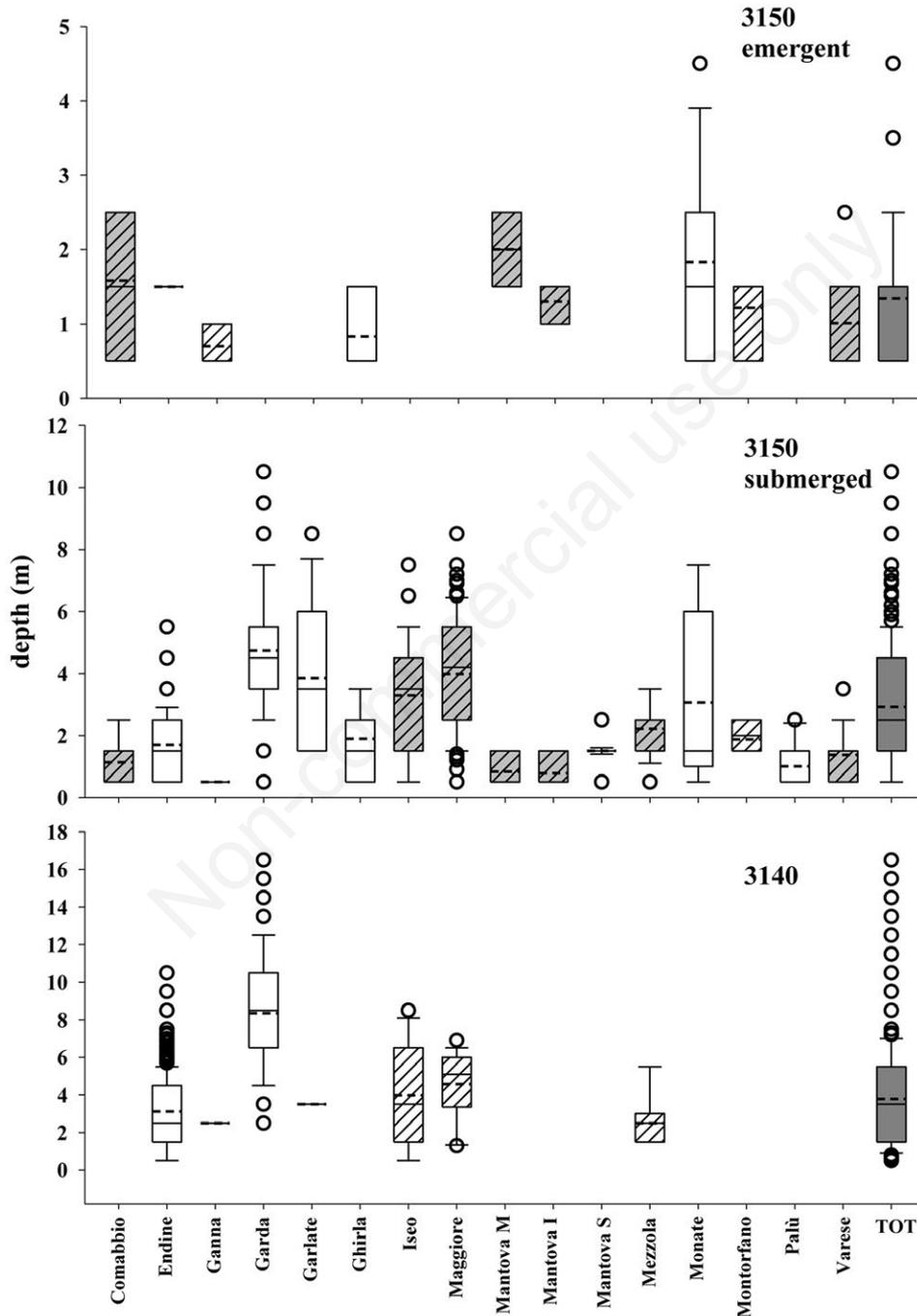
### Aquatic EU habitats in the lakes of Lombardy

On the basis of Lombardy Regional Environmental Protection Agency surveys, the results of our study highlight the local predominance of vascular submerged communities in the depth range of 1 to 4 m, which are dominated by *V. spiralis*, *N. marina* and *M. spicatum*. Indeed, vascular plant communities were limited to the upper water depths, whereas communities that live at the maximum growing depths in lakes are composed of non-vascular species (Hutchinson, 1975). Accordingly, we substantiated the existence of a fairly clear differentiation between the two habitats we analyzed (EU 3140 and 3150), thereby highlighting the pivotal role played by the ability of vascular and charophyte species to capture light in shaping the depth distribution of aquatic vegetation. In-

deed, EU habitat 3140 was found to mainly occur in deep lakes, not all of which are included in the local Natura 2000 network. This suggests that the regional policy regarding this habitat needs to be reassessed. This finding is in agreement with a preliminary evaluation of the repre-

sentativeness of *Chara*-dominated stands in Lake Garda, which may account for approximately 20% of the overall area occupied by this habitat at national scale (1000-1200 ha out of a total of ~5000 ha) (Bolpagni *et al.*, 2013).

In the lakes we investigated, EU habitats 3140 and



**Fig. 2.** Box plot graph of the depth distribution of the diagnostic species of the EU habitats 3150 (considering both the emergent and submerged plant communities) and 3140. The box indicates 25<sup>th</sup>-75<sup>th</sup> percentile; the solid and dotted lines indicate median and mean values, respectively. The lakes with previous habitat records are highlighted in grey, the coarse pattern indicates the lakes included in the Natura 2000 network. Mantova M, Mantova Middle; Mantova I, Mantova Inferior; Mantova S, Mantova Superior; TOT, all data together.

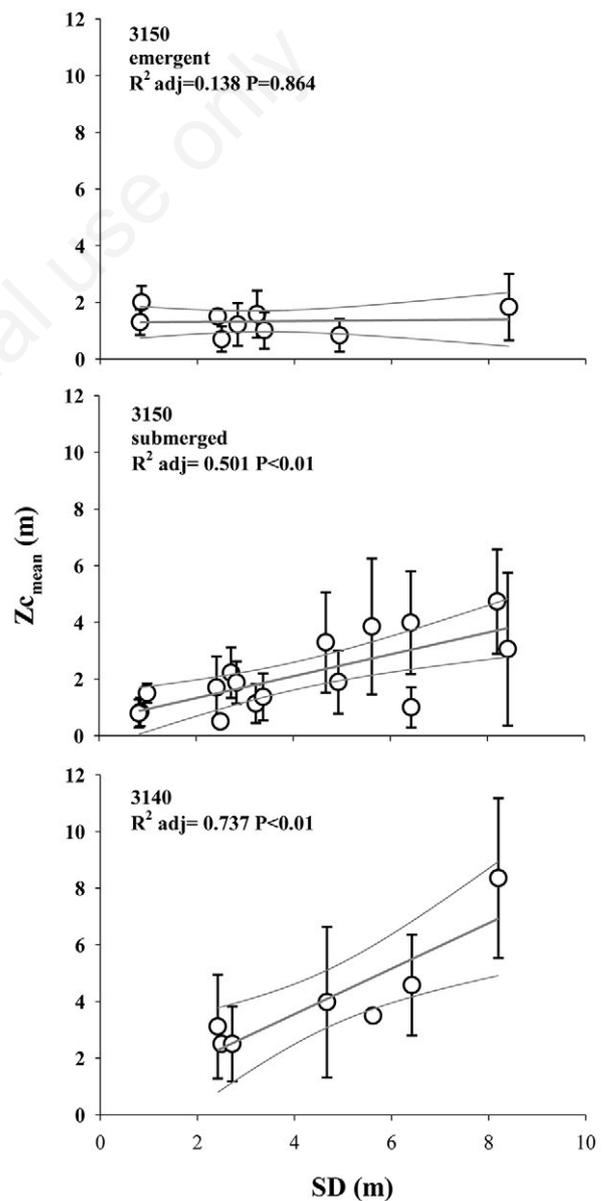
3150 displayed a comparable range of colonized depths that was partially overlapped (0.5-4.5 m vs 1.5-5.6 m; Fig. 3), which reflects the wide range of SD conditions considered. Hence, the variability in underwater light conditions affects both the presence/absence and the extent of the depth range colonized by diagnostic species, which change significantly in different lakes. Although the dataset we used was limited in size, our findings corroborate the strong interdependence between aquatic habitat depth ranges and the underwater light environment (Canfield *et al.*, 1985; Chambers and Kalff, 1985). This is substantiated by the differences that emerged between the linear regression models generated, with the increased steepness of the estimated slope revealing a greater sensitivity of charophytes to a progressive worsening in water conditions (Fig. 3). Even a small reduction in SD may be assumed to be associated with a non-negligible reduction in the area occupied by *Chara*-dominated stands. By contrast, vascular-dominated stands appear to be more “resilient”, with a lesser dependency on SD variations. One explanation for the findings yielded by the lakes we investigated may be the replacement of sensitive aquatic vascular plants (mainly species belonging to the genus *Potamogeton*) by more tolerant species (*e.g.*, *M. spicatum*, *V. spiralis*) or alien taxa, such as *Lagarosiphon major* (Ridl.) Moss and *Elodea nuttallii* (Planch.) H. St. John (Bolpagni *et al.*, 2015; Soana and Bartoli, 2013, 2014).

### Practical remarks on the WFD and HD integration

Although the standardized WFD methods used for macrophyte characterization cannot be considered exhaustive to describe vegetation, mainly because they do not provide a full and detailed floristic account of plant communities (Oggioni *et al.*, 2011), it is worth bearing in mind that their primary objective is the ecological classification of water bodies. This type of information can be used to define the spatial distribution and depth rearrangement of the diagnostic species of many EU habitats of conservation concern. Furthermore, if compared with the methods commonly used to characterize aquatic vegetation (*i.e.*, phytosociology), the standardized WFD methods ensure the acquisition of ancillary data on the physical and chemical conditions of the colonized environments. These data are important as a means of assessing the quality of habitats and colonized sites as well as of evaluating the future prospects of habitats (Collins *et al.*, 2012; Louette *et al.*, 2015 and references therein). This is a focal point that supports the integration of current programs aimed at protecting lacustrine aquatic EU habitats not only in Lombardy, but also across Europe as a whole. Bearing all this in mind, we wish to stress the need to implement strategies for the conservation of aquatic habitats by integrating trophic trends of lakes in terms of nutrient availability and water transparency as well as of the rela-

tive size of water bodies (Azzella *et al.*, 2014b).

By focusing on the distribution patterns of aquatic habitats of conservation concern, this paper attempts, to our knowledge for the first time, to actively combine the WFD directive with the HD. The standard WFD data on macrophytes appear to complement the needs and the requirements of HD, especially for defining the area (range) occupied by a specific lacustrine aquatic EU habitat. The use of data collected by the WFD will also enable us to make robust comparisons between biogeographic regions and states, thereby strengthening evaluations of current trends



**Fig. 3.** Linear regression between water transparency, expressed as Secchi disk (SD, in m), and the mean depth of the habitat distribution ( $Z_{c_{mean}} \pm$  standard deviation, in m).

in aquatic vegetation on a large spatial scale. Furthermore, the possibility of gathering reliable and valid information on the physical and chemical conditions of water and sediments may make a fundamental contribution to defining favourable conservation conditions of habitats as well as to planning the distribution of areas of aquatic vegetation in response to climate change or direct human impacts (*i.e.*, pollution, water abstraction). This is especially important for aquatic environments, which are among the most vulnerable ecosystems in the world (Verhoeven *et al.*, 2006).

Within this context, a better integration between HD, WFD and remote sensing techniques may represent a key strategy to obtain reliable information on the ranges occupied and to examine current dynamic trends in aquatic vegetation, as has previously been shown by many authors (Villa *et al.*, 2013, 2014, 2015; Giardino *et al.*, 2015). Indeed, remote sensing is an effective tool for recognizing and mapping vegetation thanks to the advantages of the synoptic view (in time and space) over the field survey (Vis *et al.*, 2003; Silva *et al.*, 2008). Furthermore, the vegetation indices yielded by airborne and satellite images can effectively be used to implement standard monitoring approaches, thereby providing new opportunities in the assessment of vegetation status, growth or biophysical features. This may be considered a further step toward maximizing monitoring efforts and assessment outcomes.

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## CONCLUSIONS

This study highlights, for the first time, the usefulness of approaches implemented and adopted by the WFD to assess macrophyte communities in lakes as a means of also evaluating the presence of EU habitats and the size of the areas they occupy. Furthermore, the data collected within the framework of the WFD monitoring networks are comparable because they are obtained by means of similar sampling procedures, calibrated at the ecoregion scale by specific Geographical Intercalibration Groups. In addition, the use of these data may minimize the limitations usually associated with local flora and vegetation surveys, which are often restricted to a single, or a few, Natura 2000 sites. WFD data effectively address the urgent need to extend the information available on the presence and state of conservation of EU habitats beyond the Natura 2000 network.

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