How do low-abundance taxa affect river biomonitoring? Exploring the response of different macroinvertebrate-based indices

Simone GUARESCHI,1* Alex LAINI,2 Maria M. SÁNCHEZ-MONTOYA1,3

1Department of Ecology and Hydrology, Regional Campus of International Importance “Campus Mare Nostrum”, University of Murcia, Espinardo Campus, 30100 Murcia, Spain; 2Department of Life Science, University of Parma, Viale G.P. Usberti 33/A, 43124 Parma, Italy; 3Leibniz-Institute of Freshwater Ecology and Inland Fisheries (IGB), Müggelseedamm 310, 12587 Berlin, Germany

*Corresponding author: simone.guareschi@um.es

ABSTRACT

The contribution of rare taxa to aquatic bioassessments remains a subject of debate, and generates contrasting positions among researchers. Very little is known about the effect of low-abundance taxa (LAT) for calculating both single and multimetric macroinvertebrate-based indices, as well as the ecological status classification. In this study, we aimed to: i) identify the aquatic macroinvertebrates that need special attention during index applications given their low abundance; ii) analyse the effect of excluding LAT on single (IBMWP and IASPT) and multimetric (STAR_ICMi) biological indices; and iii) investigate the influence of LAT on river ecological status assessments. To this end, two different river basins in SE Spain and N Italy with contrasting climatic conditions and river types were selected. Our results showed that almost all the taxa at the family level can act as low-abundance taxa. In particular, the LAT belonged mainly to Diptera, Trichoptera, Coleoptera, Gastropoda and Hemiptera. The IndVal analysis stressed Tabanidae, Cordulegasteridae and Hydroptilidae as the most characteristic low-abundance families in the Spanish data set, while Dryopidae and Athericidae were identified mostly in N Italy. Excluding LAT affected the studied index values and the resulting bioassessment classification, except for the IASPT index. Loss of the entire LAT pool reduced the ecological status for 78% of the samples for the IBMWP index. Changing took place in 41% of the samples when considering the STAR_ICM index. Relevant changes were detected even when considering loss of 50% of the LAT, especially with the IBMWP index. Similar values and patterns were obtained in each considered quality class and river type. Our results provide useful information about controversial taxa and stress the significance of LAT in river biomonitoring. Excluding LAT is discouraged, although different responses according to the considered index were detected. The IBMWP index always obtained lower values (coupled mainly with an underestimation of ecological classes), while STAR_ICMi and, especially the IASPT index, were less affected by excluding LAT. Paying special attention to all the protocol application stages is recommended, with emphasis placed when using the IBMWP index.

Key word: Bioassessment; rare taxa; ecological classes; IBMWP; IASPT; STAR_ICMi.

INTRODUCTION

River assessment is extensively based on aquatic macroinvertebrates and numerous biomonitoring indices have been developed for this purpose (Doledé and Stattzer, 2010; Birk et al., 2012; Bo et al., 2016). These indices are generally based on rapid biological assessment protocols (Barbour et al., 1999), followed to obtain a practical and rigorous assessment of river ecological conditions. However, macroinvertebrate sampling activities may reflect just one part of the community, and documenting some taxa (e.g., with either low abundances or small distribution ranges) is often time-consuming and cost-intensive in terms of fieldwork, laboratory work, data processing and analyses (Nijboer and Schmidt-Kloiber, 2004).

These taxa are generally called “rare” (Gaston 1994; Nijboer and Verdonschot, 2004) and their contribution to aquatic biomonitoring has received particular attention in multivariate analyses for bioassessment (Cao et al., 2001; Marchant 2002) and predictive methods, especially in Anglo-Saxon countries (e.g., the RIVPACS approach; Clarke and Murphy, 2006; Van Sickle et al., 2007). Nevertheless, this debate is ongoing with researchers offering contrasting positions and findings (Cao and Williams, 1999; Marchant et al., 1999; Poos and Jackson, 2012). Very little is known about the effect of rare taxa for calculating single or multimetric macroinvertebrate-based indices or about the effect on ecological status assessments, especially in South Europe.

Contrary to taxa with small distribution ranges, low-abundance taxa (herein referred to as LAT) might be present in numerous samples, sites and habitats, but have the general peculiarity of presenting low number of individuals during sampling activities. Although excluding this kind of taxa may be cost-effective (especially in the field) when applying rapid biological assessment protocols, LAT may be common contributors to the assemblage structure in undisturbed streams (Robinson et al., 2000), and might even represent the largest overall richness component (Cao et al., 1998).

The low abundance of a macroinvertebrate taxon, or
even its absence, from a sample can be attributed to both structural (i.e., an inhospitable environment that does not allow a species to establish) and stochastic (i.e., incomplete characterisation of a macroinvertebrate community) causes (Gray, 2005). In this study, we focused specially on this second aspect as community characterisation can be strictly related with difficulties during fieldwork or laboratory procedures, such as operator efficiency (Metzeling et al., 2003), sorting errors (Haase et al., 2010), or subsampling methods (Nichols and Norris, 2006). As a result, some methods and procedures may be biased towards large, abundant and widely distributed taxa (Gillies et al., 2009; Haase et al., 2010). LAT have fewer chances of being sampled, so their distribution range can be underestimated (Nijboer and Verdonschot, 2004). The imperfect detection of these taxa can be a common problem that may affect numerous biological metrics; for instance, total taxon richness and EPT richness, which are two of the main metrics used in biomonitoring (Birk et al., 2012).

In order to improve knowledge on this topic and to provide applied information and recommendations for environmental agencies, technicians and researchers, we aimed to: i) identify the aquatic macroinvertebrates that need special attention while applying protocols by discussing their ecological preferences and features. When dealing with the other two, we investigated whether the exclusion of these taxa could be considered crucial for determining not only the index final values, but also ecological classes, to provide useful information about efficient future sampling strategies or line guides. In this way, the comparison of different index outcomes and behaviours is possible and useful as study metrics and indices are among the most widely used tools in river bioassessments (Armitage et al., 1983; Munné and Prat, 2009; Laini et al., 2014; Buss et al., 2015). Finally, controversial issues and possible future implications are discussed.

METHODS

Study area and sampling sites

The present study was carried out by analysing different biogeographical and climatic regions (S Spain and N Italy). We focused on the Segura Basin (SE Spain, Fig. 1a) with 30 sites located in the Murcia, Andalusia and Castilla-la Mancha regions, which were sampled mainly during three seasons in 2000-2003 for 81 sample data (Supplementary Tab. 1). The Segura Basin is an environmentally diverse basin, considered a good candidate to be utilised as a Mediterranean pilot basin (Bruno et al., 2014). Agricultural impacts and dam regulation have been recognised as representing the most important pressures on aquatic ecosystems in this catchment (Kroll et al., 2013; Sánchez-Montoya et al., 2009; Bruno et al., 2014). We considered types, reference values and class boundary values according to those proposed by Sánchez-Montoya et al. (2007, 2010) and developed in the Spanish study area. These biological data and types were developed as part of the GUADALMED II Project for Mediterranean Spanish rivers (Prat, 2004). To obtain a representative data set, the sites of four of the five river types in Mediterranean rivers were considered (including temporary and perennial, and different lithologies; Tab. 1).

We also studied the effect of LAT on biomonitoring in the Po Basin (N Italy, Fig. 1b). It is the widest river basin...
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in Italy. We considered 29 representative sites (Lombardia and Emilia-Romagna regions) that were sampled mainly twice (86 sample data; Supplementary Tab. 1). The macroinvertebrate data originate from the fieldwork activities performed from 2009 to 2015. Similar to the Segura Basin, agriculture and flow alterations can be considered the main pressures in the area (Laini et al., 2011; Guareschi et al., 2014). We considered types, reference values and class boundary values according to those published in the Italian ministerial decree DM 260/2010. The data set encompasses three river types from lowland to Apennine streams (Tab. 1, see details on Italian hydro-regions in Buffagni et al., 2008).

According to the Köppen-Geiger climate map of Europe (Peel et al., 2007), the study area in SE Spain is classified mostly as arid (code: BSk), while Italian regions are temperate (code: Cfa).

Macroinvertebrate indices: IBMWP, IASPT and STAR_ICMi

We defined low-abundance taxa as all the taxa whose abundance was $\leq 3$ individuals per sample, which corresponds to the first abundance class in the IBMWP index, and as $\leq 3$ per 0.5 m$^2$ for the quantitative STAR_ICMi. Similar criteria have been considered in other studies (Bradley and Ormerod, 2002; Gillies et al., 2009). The response of three different macroinvertebrate-based indices, IBMWP (Iberian Biological Monitoring Working Party), IASPT (IBMWP value/number of families) and STAR_ICMi (Intercalibration Multimetric Index), to LAT exclusion was tested. The first two indices were tested with the Spanish data set, while the Italian data set was used with the last one (see the details below).

The taxonomic resolution needed to calculate these indices is the family level and five ‘Ecological status’ levels have been established according to European legislation (WFD 2000/60/CE, European Commission 2000) for IBMWP and STAR_ICMi. IBMWP is the most widely used index in Spanish Mediterranean rivers (Alba-Tercedor et al., 2002; Munné and Prat, 2009). It is currently the official index proposed in Spanish rivers (MAGRAMA, 2015) and is an adaptation of the British BMWP scoring system for the Iberian Peninsula, where each family presents a score from 0 to 10 according to their known tolerance to pollution. The IBMWP index is a single metric index (Munné and Prat, 2009) and the sample value is obtained by summing these family scores. This index is considered a simple one (Couto-Mendoza et al., 2015) with a multi-habitat semi-quantitative kick procedure in the field ( Jáimez-Cuéllar et al., 2002). IASPT represents a sensitive taxa index that is simply calculated for each sample as the IBMWP value divided by the number of scoring families detected. It was calculated only for the Spanish data set in this study as this index is currently recognised as a biomonitoring tool in Spain and is commonly used in this area (e.g., Sánchez-Montoya et al., 2010). Unlike the other indices, STAR_ICMi is a multimetric index and is the official index used in Italy for assessing water course quality statuses according to European legislation (Buffagni et al., 2006, 2008). It requires a quantitative sampling activity referee on a surface (0.5 m$^2$ or 1.0 m$^2$ depending on the Italian hydro-regions) and has also been used in Europe as an Intercalibration Common Metric Index (Buffagni et al., 2006). It is composed of six metrics: ASPT (Average Score Per Taxon), logarithm of the abundances of the selected families of Ephemeroptera, Plecoptera, Trichoptera and Diptera (1-GOLD), number of LAT were calculated. To refine the analysis and to deal with the first goal, the Indicator Value analysis (Ind-Val) was carried out to select specific rare families per river type (Dufrêne and Legendre, 1997; De Cáceres et al., 2010). Such affinity was calculated according to the frequency of each taxon in the previously identified groups by taking into account only the sub-communities that composed of LAT. The significance of the Indicator Value (IV) was tested by a Monte Carlo test (999 runs), and the alpha level was set at 0.05.

In order to test the effect of excluding LAT on river biomonitoring indices, they were removed from each

**Statistical analyses**

First of all, the percentage of LAT and the contribution of each taxonomic group (mainly Order) to the total number of LAT were calculated. To refine the analysis and to deal with the first goal, the Indicator Value analysis (Ind-Val) was carried out to select specific rare families per river type (Dufrêne and Legendre, 1997; De Cáceres et al., 2010). Such affinity was calculated according to the frequency of each taxon in the previously identified groups by taking into account only the sub-communities that composed of LAT. The significance of the Indicator Value (IV) was tested by a Monte Carlo test (999 runs), and the alpha level was set at 0.05.

In order to test the effect of excluding LAT on river biomonitoring indices, they were removed from each

<table>
<thead>
<tr>
<th>Code</th>
<th>Description</th>
</tr>
</thead>
<tbody>
<tr>
<td>T1</td>
<td>Temporary streams</td>
</tr>
<tr>
<td>T2</td>
<td>Evaporite calcareous at medium altitude</td>
</tr>
<tr>
<td>T4</td>
<td>Calcareous headwaters at medium and high altitude</td>
</tr>
<tr>
<td>T5</td>
<td>Large watercourse</td>
</tr>
<tr>
<td></td>
<td><strong>Italian dataset</strong></td>
</tr>
<tr>
<td>Code</td>
<td>Description</td>
</tr>
<tr>
<td>------</td>
<td>-----------------------------------</td>
</tr>
<tr>
<td>06GL</td>
<td>Floodplain watercourse originated from lakes</td>
</tr>
<tr>
<td>06SS</td>
<td>Floodplain streams</td>
</tr>
<tr>
<td>10SS</td>
<td>Northern Apennine streams</td>
</tr>
</tbody>
</table>

**Tab. 1.** Code and description for each river type in the study area. The details for each site are available in Supplementary Tab. 1.
sample after considering 100% and 50% of LAT (see details below). Then each index was recalculated and all the samples were reclassified in the corresponding ecological classes. The first case corresponded to excluding the entire pool of LAT (e.g., relevant problems in the sorting or picking phase, operators have little experience). We also tested a more conservative exclusion threshold, which corresponded to a 50% loss of LAT. To do this, the excluded LAT were obtained by considering 100 randomisations of LAT constant loss one by one, and by stopping at the value that equalled a 50% loss of the total LAT pool. At this point, we once again recalculated each index and all the samples were reclassified in the corresponding ecological classes.

The differences between the index values obtained before and after totally excluding LAT were statistically tested by a non-parametric Mann-Whitney test. The results were reported for both the Spanish and the Italian data sets, and even after splitting each data set according to its ecological classes and river type. The behaviour of each index was also studied graphically by focusing on the trend obtained after performing the LAT loss simulations.

To determine whether possible changes in ecological status were equally distributed among the ecological quality classes, we classified the sample data into three classes: High, Good and Less than Good (codes: H, G, < G). The boundary between Good and Moderate is crucial according to European legislation (WFD 2000/60/CE) because it sets the targets for restoration plans in measuring programmes of water bodies which fail the environmental objectives of achieving a good ecological status. All the statistical analyses were performed with the statistical computing R software (R Development Core Team, 2013) with packages “Vegan”, “ade4” and “indicspecies”.

RESULTS

Low-abundance taxa characterisation

Ninety-five macroinvertebrate taxa (92 families, plus Hydrachnidia, Ostracoda, Oligochaeta) were found in the Spanish basin (mean±SD = 30±11; SD, standard deviation), with a mean value of 11 taxa (±5) as LAT (details in Tab. 2). Ninety-one of the 95 taxa were identified as rare at least once. The taxa with low abundances belonged mainly to Diptera (20%), Trichoptera and Coleoptera (both 14%) in this area (Fig. 2).

The IndVal analysis stressed just one indicator family in river types T1 and T2 as Baetidae and Tabanidae, respectively. Larger groups of indicator families were obtained for types T4 and T5. Cordulegasteridae and Gyrinidae showed the highest IV for T4, with Hydroptil-

| Tab. 2. Mean number (±SD) of the low-abundance taxa (LAT) for each river type in each data set. The total number of taxa (TT) and the total number of LAT are also displayed. |
|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|
| T1 | T2 | T4 | T5 | Total | 06GL | 10SS | 06SS | Total |
| LAT | 8±5 | 10±3 | 15±3 | 10±5 | 91 | 8±3 | 10±4 | 8±3 | 90 |
| TT | 73 | 73 | 83 | 55 | 95 | 68 | 66 | 42 | 92 |

Fig. 2. Distribution of LAT (in %) among the main taxonomic groups (mainly Order). Spanish results are labelled in grey and Italian ones in the darker tone.
idiae, Ceratopogonidae and Philopotamidae in the T5 ecosystems (Tab. 3).

Regarding the Italian dataset, 92 taxa (91 families, plus Hydrachnidia) were detected (21±5 per sample) with a mean value of 9±3 LAT per sample. Ninety taxa were identified as rare at least once (Tab. 2), where the highest percentage of low-abundance taxa belonged to Trichoptera (16%), followed by Diptera, Gastropoda and Coleoptera (Fig. 2).

The IndVal analysis showed only two families as being characteristic LAT in the 06GL river type (Corbiculidae and Aphelocheiridae), whereas a larger number of taxa were found in the other types (Tab. 4). Dryopidae, Leuctridae and Oligoneuriidae showed the highest IV for...
floodplain rivers (06SS), while Athericidae and Empididae were identified in Apennine streams (10SS, Tab. 4).

**Effect of excluding low-abundance taxa on single metric indices: IBWMP and IASPT**

Excluding the entire pool of LAT (100%) resulted in significant differences compared with the IBMWP results (W=3321; P < 0.0001). The original IBMWP mean value was 148; on the contrary when the index was recalculated, we obtained a mean value that equalled 93. The ecological quality classification of 29 of the 30 sampling sites changed at least once, and the ecological status of 78% of the samples changed to a minor class (see Tab. 5). Moreover, 82% of the values first classified as High shifted to a minor status (Good). Similar results were obtained (74-75%) when the starting point was Good or Less than Good. Substantial changes were noted for each river type in the analysis (from 33% to 100%), which was especially true for the T2 streams that exhibited huge changes in quality assessment when the starting points were High or Less than Good classes.

The 50% randomised LAT loss also led to relevant changes in the IBMWP values, but the percentage of class changes lowered compared to the total LAT loss (Tab. 5). The ecological status of about half the overall samples (48%) changed (24 of the 30 sites changed at least once). Once again, percentages were similar among different classes (43-53%). When focusing on each river type separately, the patterns were basically the same (compared with the total LAT loss), with changes falling between 33% and 100% of cases depending on ecological class. When LAT were not considered, all the river types presented similar patterns, and their IBMWP values clearly and constantly dropped (Fig. 3).

Unlike the IBMWP results, the IASPT recalculations (when all the LAT were excluded) gave values that did not statistically differ from the original ones (W=1382, P=0.1906). Graphically, the IASPT behaviour of LAT loss appeared constant and displayed a horizontal trend, with the only exception being river type T1, where more diversified responses were obtained (Fig. 4).

**Effect of excluding low-abundance taxa on a multimetric index: STAR_ICMi**

The exclusion of all the LAT (100%) led to significant differences when we compared the STAR_ICMi results (W=3781; P < 0.0001). The original STAR_ICMi mean value was 0.74, which became 0.62 when recalculated. The ecological quality classification of 19 of the 29 sampling sites underwent at least one class change. The ecological status of 41% of all the samples changed to a minor class. Similar values and patterns were obtained through different ecological classes (36-41%), except when the starting point was High class (100% change; but a limited number of data were available; Tab. 6).

The 50% randomised loss of LAT brought about changes in the STAR_ICMi values (Tab. 6). The ecological status of 33% of the overall samples changed (17 of 29 sites underwent at least one change). Once again, the percentages were similar among the different classes (23-36%), expect when the original starting point was the High class. When focusing on each river type separately, and compared to the total LAT loss, the percentage of class changes lowered, except for river type 06SS (Tab. 6). When we graphically analysed the behaviour of STAR_ICMi, the response patterns were not as clear as they were for the other indices. A decreasing tendency seemed the commonest behaviour in the three studied river types when LAT were excluded (Fig. 5).

**DISCUSSION**

**Low-abundance taxa**

Despite the climatic and environmental differences in the two study areas, LAT belonged mainly to the same or-

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Tab. 5. Number of changes of ecological quality classes for the 100% or 50% LAT loss. The results are displayed after considering the overall number of samples (Total), divided by river type (T1, T2, T4, T5) and ecological class (the classification represents the starting point class before excluding LAT) in the Spanish data set. Number of sampling data per river type is also displayed.

<table>
<thead>
<tr>
<th>Spain (Segura Basin)</th>
<th>Total</th>
<th>T1 (n=16)</th>
<th>T2 (n=25)</th>
<th>T4 (n=34)</th>
<th>T5 (n=6)</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>100%</td>
<td>50%</td>
<td>100%</td>
<td>50%</td>
<td>100%</td>
</tr>
<tr>
<td>Changes</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>No changes</td>
<td>78%</td>
<td>48%</td>
<td>12 (75%)</td>
<td>9 (56%)</td>
<td>21 (84%)</td>
</tr>
<tr>
<td></td>
<td>22%</td>
<td>52%</td>
<td>4 (25%)</td>
<td>7 (44%)</td>
<td>4 (16%)</td>
</tr>
<tr>
<td>Classes</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>High</td>
<td>82%</td>
<td>53%</td>
<td>2 (67%)</td>
<td>2 (67%)</td>
<td>5 (100%)</td>
</tr>
<tr>
<td>Good</td>
<td>74%</td>
<td>43%</td>
<td>7 (80%)</td>
<td>5 (63%)</td>
<td>13 (76%)</td>
</tr>
<tr>
<td>&lt;Good</td>
<td>75%</td>
<td>50%</td>
<td>3 (60%)</td>
<td>2 (40%)</td>
<td>3 (100%)</td>
</tr>
</tbody>
</table>
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ders in both data sets (e.g., Diptera, Trichoptera and Coleoptera) and were basically those stressed by Nijboer and Schmidt-Kloiber (2004) for Dutch streams. When focusing on the Spanish data set, and according to the criterion of IV>25 as a key value to judge an indicator taxon as being adequate (Bonada et al., 2008, following Dufrène and Legendre, 1997), Baetidae was not a good indicator for the T1 streams. On the contrary in the T2 streams, Ta-

![Graph showing IBMWP index behaviour per river type (T1, T2, T4, T5) after randomisations of constant LAT loss.](image)

**Fig. 3.** IBMWP index behaviour per river type (T1, T2, T4, T5) after randomisations of constant LAT loss. The boundary value between the Good and Moderate conditions is displayed. The index values are displayed on the y-axis and the number of considered LAT is found on the x-axis.

**Tab. 6.** Number of changes of ecological quality classes for the 100% or 50% LAT loss. The results are displayed after considering the overall number of samples (Total), divided by river type (06GL, 10SS, 06SS) and ecological class (the classification represents the starting point class before excluding LAT) in the Italian data set. Number of sampling data per river type is also displayed.

<table>
<thead>
<tr>
<th>Classes</th>
<th>Total 100%</th>
<th>Total 50%</th>
<th>Italy (Po Basin)</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>06GL (n=55)</td>
<td>10SS (n=24)</td>
<td>06SS (n=7)</td>
</tr>
<tr>
<td>Changes</td>
<td>41%</td>
<td>33%</td>
<td>21 (38%)</td>
</tr>
<tr>
<td>No changes</td>
<td>59%</td>
<td>67%</td>
<td>34 (62%)</td>
</tr>
<tr>
<td>Classes</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>High</td>
<td>100%</td>
<td>100%</td>
<td>-</td>
</tr>
<tr>
<td>Good</td>
<td>41%</td>
<td>36%</td>
<td>9 (38%)</td>
</tr>
<tr>
<td>&lt;Good</td>
<td>36%</td>
<td>23%</td>
<td>12 (39%)</td>
</tr>
</tbody>
</table>
banidae needed special attention during sample and sorting activities, as did a large group of taxa dominated by Cordulegasteridae and Hydroptilidae for the T4 and T5 streams, respectively. Cordulegasteridae larvae (genus *Cordulegaster*), Tabanidae and Athericidae (see below) are generally predators of other aquatic invertebrates (Tachet *et al.*, 2010). Members of these feeding groups have already been stressed as being relatively less abundant than prey species (Spencer, 2000). Hydroptilidae larvae (Trichoptera) usually prefer specific microhabitats with submerged vegetation where they anchor their cocoons which, coupled with their limited size and lack of movements (under dry and wet conditions) (Tachet *et al.*, 2010), may make them particularly difficult to detect.

None of the taxa stressed by the analysis was characterised by extreme IBMWP scores, and higher values were depicted by some Odonata (*e.g.*, Cordulegasteridae) or Trichoptera (*e.g.*, Philopotamidae). Most of the indicator taxa presented intermediate values, basically with scores of around 3, 4, or even 6 points, as confirmed by the IASPT response to LAT loss (Fig. 4). The mean taxa scores did not change, but stabilised with values between 4 and 6. In this situation, LAT presented a comparable IBMWP score with the commonest taxa. These results (Fig. 4) agree with the research of Nijboer and Schmidt-Kloiber (2004), who found that the mean scores for saprobic valences were similar for both taxa types (low and high abundances). If LAT had been taxa with an extreme IBMWP score (1 or 10), the IASPT responses to their loss would have shifted towards a marked change in their slopes.

According to the Italian results, it was difficult to suggest specific LAT for the 06GL rivers if we considered that the IndVal stressed taxa with IV came close to 25. On the contrary, Dryopidae (Coleoptera) and Athericidae (Diptera) were underlined as characteristic LAT in 06SS and 10SS, respectively. Special attention needs to be paid

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**Fig. 4.** IASPT index behaviour per river type (T1, T2, T4, T5) after randomisations of constant LAT loss. The index values are displayed on the y-axis and the number of considered LAT is found on the x-axis.
to these taxa in all the sampling phases and for all laboratory activities. Dryopidae larvae are generally semi-aquatic or riparian (Jách and Balke, 2008) and are, therefore, rarer to detect during sampling activity as only adults are strictly aquatic. Furthermore, Dryopidae have already been stressed as infrequent in low mountainous streams in other European countries (Slovakia: Zatovičová et al., 2004). Regarding specific LAT features, other studies have generally found that these taxa are associated with standing or slowly flowing waters (Nijboer and Schmidt-Kloiber, 2004).

Our results in both data sets only partially confirmed this finding by considering that a heterogeneous group of LAT was found. This group was composed of some lotic preference taxa, but also some highly diverse taxa (Coleoptera, Diptera), and even some strictly lentic ones (Odonata, Gastropoda, and Hemiptera). However, benthic sampling methodologies like those utilised (D-net and Surber, following official legislation) could also be the reason for the low abundance found in some groups (e.g., swimmers or surface skaters). Most of these taxa (e.g., Coleoptera or Hemiptera) live on the water surface or among vegetation, and can easily escape sampling nets. Nevertheless, the LAT identified herein cannot be considered rare or endemic in terms of small range areas, especially given taxonomic resolution (family level).

**Performance of indices and implication for bioassessment**

Excluding LAT resulted in underestimating ecological status (values lower than reality) for STAR_ICM and IBMWP indices. This effect became considerably stronger after applying the IBMWP index, with relevant

![Fig. 5. STAR_ICM index behaviour per river type (10SS, 06GL, 06SS) after randomisations of constant LAT loss. The boundary value between the Good and Moderate conditions is displayed. The index values are displayed on the y-axis and the number of considered LAT is found on the x-axis.](image-url)
changes in the ecological evaluations. This is strictly related with the index formula, which exclusively represents a sum of family scores and is strongly affected by a reduction in the total recognised taxa.

Different outcomes were detected when focusing on the IASPT index. This index did not appear sensitive when LAT were excluded from three of the four Mediterranean river types, and displayed a contrasting response in temporary rivers (T1). Applying biomonitoring procedures to temporary aquatic ecosystems is currently one on the main bioassessment challenges and requires specific tools (Nikolaidis et al., 2013; Datry et al., 2014; Prat et al., 2014). The general observed lack of sensitivity of this index can be considered an interesting attribute (i.e., no specific knowledge or experience required), but this approach may be questionable, or even dangerous (Metzeling et al., 2003; Haase et al., 2006). We should also consider that the IASPT index has been reported to be less sensitive to stressors than the IBMWP index in Mediterranean rivers (Sánchez-Montoya et al., 2010).

The Italian index STAR_ICM gave an intermediate response compared to the other indices. Following LAT loss, the results were generally underestimated, but more slightly than for the IBMWP index. Once again, this is probably related with the index formula. As it was a multimetric index, and not just a mathematic sum, it could be less affected by the reduction in the total taxa considered herein. Richness metrics formed part of the index (e.g., total of families, total EPT taxa), but the metrics with the heaviest weight in the definition of the index was the Average Score Per Taxon (ASPT) value. Knowing the specific behaviour of the Spanish ASPT (investigated herein) could help explain the attenuated STAR tendencies.

Unlike our results, Nijboer and Schmidt-Kloiber (2004) reported that excluding LAT overestimated the ecological quality class in Dutch lowland streams with the AQEM software (Hering et al., 2004). These contrasting findings, which are especially strong between IBMWP and AQEM, may be due to the different geographic contexts and river types considered (e.g., Northern vs Southern Europe), but could also be due to specific differences between protocols (e.g., taxonomic resolution, sampling area). Despite the AQEM method and STAR_ICMi presenting similar procedures (Hering et al., 2006), contrasting results were obtained as LAT exclusion did not generally lead to overestimates in the Italian data set. Nevertheless, in both cases (underestimation and overestimation), these mismatches could have serious consequences for environmental agencies and water managers during environmental management (e.g., restoration where it is not necessary), which can provide equivocal pictures of river ecosystem health.

CONCLUSIONS

Our study is one of the first attempts to test the effects of excluding LAT in several widely used European indices. Previous studies have tested macroinvertebrates at the genus or species level (Cao et al., 1998; Resh et al., 2005), and have mainly examined the effects of LAT on richness metrics. Our results demonstrated that even at the family level and in two different study areas, LAT represent a large part of overall richness. Their exclusion strongly lowered the number of taxa in the samples regardless river typology, with effects on bioassessment depending on the index employed. In this context, the implementation of quality control mechanisms into macroinvertebrate assessment procedures seems recommendable.

Finally, by considering the increasing impacts and challenges of alien species on aquatic ecosystems (Havel et al., 2015; Fenoglio et al., 2016), alien taxa must be specially considered because they can be generally characterised by initial low abundances in their first invasion phases. Indeed, in our study, Corbiculidae (e.g., Corbicula fluminalis) are reported as LAT in some Italian rivers. This fact may lead to some mismatches or variations in the final index scores, and similar problems have already been stressed in other European countries (Gabriel et al., 2005). Specific reflections and future adjustments to studied systems in both countries should be made (e.g., definition of specific tolerance classes or periodically reviewing the taxa list by considering possible taxonomic modifications and future invaders) to improve ecological assessment tools in freshwater ecosystems.

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